



Terrestrial emigration behaviour of two invasive crayfish species

John Rhidian Thomas^{a,*}, Stephanie Masefield^b, Rhiannon Hunt^a, Matt J. Wood^b, Adam G. Hart^b, Jane Hallam^c, Siân W. Griffiths^a, Jo Cable^a

^a Sir Martin Evans Building, School of Biosciences, Cardiff University, Museum Avenue, Cardiff, CF10 3AX, UK

^b School of Natural and Social Sciences, University of Gloucestershire, Francis Close Hall, Cheltenham, GL50 4AZ, UK

^c School of Biological and Chemical Sciences, Queen Mary University of London, Mile End Rd, London, E1 4NS, UK

ARTICLE INFO

Keywords:

Emigration
Invasive species
Pacifastacus leniusculus
Procambarus clarkii
Overland movement

ABSTRACT

To disperse between isolated waterbodies, freshwater organisms must often cross terrestrial barriers and many freshwater animals that are incapable of flight must rely on transport via flooding events, other animals or anthropogenic activity. Decapods such as crayfish, on the other hand, can disperse to nearby waterbodies by walking on land, a behaviour that has facilitated the spread of invasive species. Overland movement could play a key role in the management of non-native crayfish, though to what extent terrestrial emigration occurs in different species is poorly understood. Here, we directly compared the terrestrial emigration tendency of two non-native crayfish species in Great Britain; red swamp (*Procambarus clarkii*) and signal (*Pacifastacus leniusculus*) crayfish. We found that both species emigrated from the water and that there was no significant difference in terms of their terrestrial emigration tendency, suggesting that there is a risk both of these species will migrate overland and disperse to new habitats. This study shows that terrestrial emigration is an important behavioural trait to consider when preventing the escape of crayfish from aquaculture and further spread of invasive species.

1. Introduction

Non-native species are common in freshwaters as humans have historically exploited these ecosystems for fishing, aquaculture and recreation, leading to the introduction of mammals, fish and invertebrates (Hulme et al., 2008; Strayer and Dudgeon, 2010; Strayer, 2010). Dispersal studies of aquatic species tend to focus on movement within a waterbody, particularly downstream migration (e.g. Bubb et al., 2002, 2004; Barson et al., 2009). The majority of introduced freshwater species must reach new habitats either by hitch-hiking on other animals or via human-mediated translocations (Shurin and Havel, 2002; Anastácio et al., 2014) to become invasive. Dispersal is a three-stage process involving emigration, inter-patch movement and immigration. Factors such as population density and competition can drive emigration (Enfjäll and Leimar, 2005; Hudina et al., 2014), though intrinsic differences in emigration tendency also exist amongst individuals, populations and species (Roland et al., 2000; Bowler and Benton, 2005; Cote et al., 2010) and invasive species are generally considered to be better dispersers compared to native congeners due to higher levels of activity (Bubb et al., 2006) and boldness (Rehage and Sih, 2004). Differences in the tendencies of non-native species to disperse, however, are less clear, though it is assumed that widespread

species have a higher dispersal tendency.

Crayfish are commercially important freshwater crustaceans that are particularly pernicious invaders in some locations (Peay et al., 2010; Gherardi, 2010). North American crayfish such as the red swamp (*Procambarus clarkii* Girard 1852) and signal crayfish (*Pacifastacus leniusculus* Dana 1852) are widely harvested species that have often escaped from aquaculture and established non-native populations (Holdich et al., 2009). The introduction of North American crayfish to Europe has led to the extirpation of native European species due to competitive displacement (Hill and Lodge, 1999; Bubb et al., 2006; Hudina et al., 2011; Hanshew and Garcia, 2012) and through the spread of the highly virulent crayfish plague, *Aphanomyces astaci* Schikora 1906 (see Holdich et al., 2014). In water, red swamp crayfish can disperse at a rate of up to 4 km in a single day (Gherardi and Barbaresi, 2000) though movement rates of between 0.61–38 m day⁻¹ are more commonly reported (Gherardi et al., 2000, 2002; Bubb et al., 2004; Aquiloni et al., 2005; Anastácio et al., 2015). Dispersal rates for signal crayfish in water are slightly lower, between 4.1 and 17.5 m day⁻¹ (Bubb et al., 2004, 2006; Anastácio et al., 2015).

As well as their potential to rapidly disperse in water, the success of some invasive crayfish can, at least partially, be attributed to their ability to colonise nearby waterbodies, navigate barriers (weirs or falls)

* Corresponding author.

E-mail address: thomasjr6@cardiff.ac.uk (J.R. Thomas).

and escape from captivity by terrestrial emigration and overland dispersal (Kerby et al., 2005; Larson et al., 2009; Holdich et al., 2014; Puky, 2014; Ramalho and Anastácio, 2015). Some crayfish, particularly burrowing species, can survive for several months out of water in burrows (Huner and Lindqvist, 1995; Kouba et al., 2016) though all crayfish can tolerate some degree of terrestrial exposure. Of the nine non-native crayfish species in Great Britain, red swamp and signal crayfish are most frequently reported dispersing overland, at least in other parts of the world (Holdich et al., 2014; Ramalho and Anastácio, 2015; Piersanti et al., 2018). Red swamp crayfish can move up to 1 km (Lutz and Wolters, 1999; Souty-Grosset et al., 2016) at a speed of 90 m h⁻¹ (Ramalho and Anastácio, 2015). We are not aware of any direct study of signal crayfish overland movement, though Holdich et al. (2014) report that signal crayfish can ‘survive for days to months’ out of water.

Overall, there is a lack of information on crayfish terrestrial emigration, largely because this behaviour is only rarely directly observed, recorded or quantified in the field. A direct comparison of invasive red swamp and signal crayfish terrestrial emigration tendencies will provide important information for predicting their future potential range expansion in Britain. Here, we tested the hypothesis that British populations of red swamp crayfish have a higher terrestrial emigration tendency and are faster when walking on land compared to signal crayfish.

2. Materials and methods

Crayfish were trapped in spring 2016, from a small private pond (Powys, Wales; signal crayfish) and from public recreational ponds (Hampstead Heath, London; red swamp crayfish). Both species were caught using cylindrical crayfish traps (‘Trappy Traps’, Collins Nets Ltd., Dorset, UK) baited overnight with cat food and transported to the Cardiff University aquarium facility, where they were maintained in a climate-controlled room set at 13 ± 1 °C, 60% RH and a 12 h light: 12 h dark cycle.

All crayfish were housed in single-species holding aquaria (density of 10 individuals per m²) filled with dechlorinated water, 2 cm gravel substrate and an excess of refugia (plant pots and PVC tubes) with no terrestrial access. Holding aquarium water was biologically filtered and a 50% water replacement was performed weekly to maintain water quality. All crayfish were fed *ad libitum* on a mix of frozen *Tubifex* bloodworm (Shirley Aquatics, Solihull, West Midlands, U.K.) and frozen peas.

The crayfish were marked using a non-toxic yellow marker (Dykem, USA) on the carapace to allow visual identification during video analysis. Furthermore, to allow individual identification if crayfish lost the mark or moulted, all crayfish were individually tagged following Bubb et al. (2002) using Passive Integrated Transponder (PIT) tags (7.5 mm PIT-tags, ISO 11,784 certified, Loligo Systems, Denmark).

2.1. Experimental design

To examine crayfish emigration tendency, an experimental arena was constructed consisting of two aquaria (L100 cm x W48 cm x H53 cm) with moveable ramps (L43 cm x W29 cm; 30° incline) that provided access to a terrestrial bridge (L240 cm x W20 cm x H20 cm)

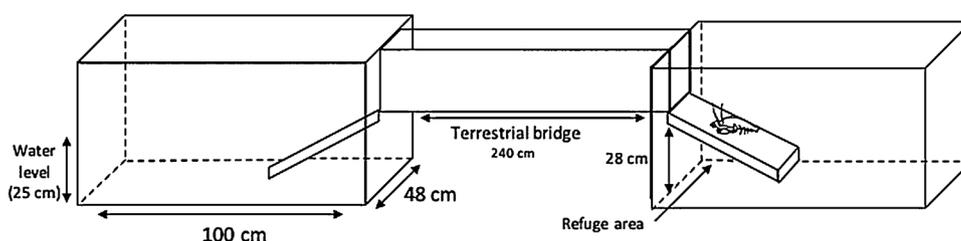


Fig. 1. Experimental arena for assessing terrestrial emigration behaviour. Crayfish could access the bridge (240 cm) using ramps, which also acted as a refuge. The water was filled to approx. 3 cm below the level of the terrestrial bridge.

(Fig. 1). Both the ramps and terrestrial bridge were coated with pea gravel to allow sufficient grip for crayfish climbing out of the water and the area under each ramp provided a shared refuge. A pea gravel bed (2 cm layer) was also provided in the aquaria. An infrared CCTV camera system (Sentient Pro HDA DVR 8 Channel CCTV, Maplin) was suspended above the arena to monitor crayfish behaviour in all experiments. All crayfish were sexed and measured (carapace length) at the start of the experiment. Signal crayfish (n = 15; 5 males and 10 females) ranged from 38.6 to 59.3 mm (mean 47.9 mm) in carapace length, whilst red swamp crayfish (n = 17; 13 males and 4 females) ranged from 47.3 to 71.3 mm (mean 58.8 mm).

The terrestrial emigration propensity and walking speed of red swamp and signal crayfish were quantified in the experimental arena (Fig. 1). At the start of each trial, individual crayfish were placed in the water on one side of the arena at 09:00 h and allowed to acclimatise until 20:00 h the same day. The lights were automatically turned off at 20:00 h and crayfish behaviour was recorded until 08:00 h the next day (12 h recording).

2.2. Ethical note

This study was undertaken in accordance with ASAB/ABS guidelines for the use of animals in teaching and research. All invasive crayfish were caught under a Natural Resources Wales licence (Trapping licence number: NT/CW065-C-652/5706/01) and held under a Cefas licence (W C ILFA 002) at Cardiff University. The crayfish were not exposed to harmful conditions during the course of the experiment, however, both species of invasive crayfish were euthanized humanely at -20 °C in accordance with the Wildlife and Countryside Act, 1981.

2.3. Statistical analyses

As crayfish are nocturnal, the 12 h observation period occurred overnight and for each individual crayfish, the number of emergences from water and time spent out of water per emergence were quantified from video recordings. A terrestrial emergence event was defined as when more than half of the crayfish body was out of the water and on the bridge. Walking speed (cm s⁻¹) was estimated over a set distance (i.e. the 240 cm bridge) and quantified only for crayfish that fully crossed the bridge.

A negative binomial GLM (Generalised Linear Model) with a log link function (using the MASS package; Venables and Ripley 2002) was used due to the zero-inflated nature of the data to determine whether crayfish species or carapace length affected the total number of times a crayfish left the water, including times they did not cross the bridge. Due to the availability of crayfish, it was not possible to test a balanced number of male and female red swamp and signal crayfish, but sex was included in the model as a nested term within species to account for potential differences.

A gaussian GLM with ‘identity’ link function was used to examine whether crayfish species or carapace length affected terrestrial walking speed (cm s⁻¹). Sex was not included in this model since all red swamp crayfish that crossed the bridge (n = 4) were male. Model selection and assumptions of normality were confirmed using residual diagnostic plots (Zuur et al., 2010). All statistical analyses were performed in R 3.4.0 (R Core Team, 2017).

3. Results

Overall, 35.3% of red swamp and 46.6% of signal crayfish left the water at least once over the 12 h nocturnal observation period. There was no significant difference in the tendency of either species to emigrate from the water nor the time they spent out of water per emergence. Certain individuals of both species tended to frequently leave the water; two red swamp crayfish left the water 17 times each, whilst two signal crayfish left the water 20 and 14 times each. The total number of emergences over the course of the experiment for both red swamp and signal crayfish was 50 and 58, respectively. Carapace length had no significant effect on terrestrial movement tendency.

Of the crayfish that emerged from the water during the 12 h observation period, red swamp and signal crayfish spent on average 6 min 42 s (SD = 124 s) and 8 min 64 s (SD = 386 s) out of water, respectively. There was no significant difference in the walking speed of either species on land, which was also unaffected by carapace length. Four male red swamp crayfish were observed to fully cross the bridge (average speed 0.703 cm s⁻¹, SD = 0.07) and six signal crayfish - two males and four females (0.601 cm s⁻¹, SD = 0.28)

4. Discussion

In the present study, terrestrial emigration occurred in both red swamp and signal crayfish from invasive British populations, with no significant difference in their tendency to leave the water or walking speed on land. These results suggest that although red swamp crayfish are generally considered to have a higher tendency to walk overland due to their burrowing tendencies and subsequent lowered risk of desiccation, signal crayfish are just as likely to walk overland, and so overland dispersal could facilitate both species' spread. In terms of their walking speed, the red swamp and signal crayfish tested here also crossed the bridge at similar speeds. Crayfish do not move particularly quickly overland, especially compared to other decapods: ghost crabs (*Ocypode* spp.), for example, can reach speeds of up to 2 m s⁻¹ (Claussen et al., 2000). Furthermore, when placed out of water, neither species of crayfish are able to direct their movements towards nearby waterbodies (Marques et al., 2015) and their walking speed decreases as they become dehydrated (Claussen et al., 2000). As such, crayfish are at a significant risk of desiccation and need to cross terrestrial barriers as quickly as possible, though neither species tested here appeared to have an advantage over the other.

Despite being the most widespread crayfish species globally, the red swamp crayfish has not yet spread far in Great Britain, potentially due to sub-optimal climate conditions (Ellis et al., 2012). Most established populations are currently found in ponds, canals and rivers around London, having first been discovered at Hampstead Heath in 1991 (Ellis et al., 2012; James et al., 2014). Signal crayfish, on the other hand, are by far the most abundant species in Britain, largely because they were introduced earlier (1970s) and on a larger scale for aquaculture (James et al., 2014; Holdich et al., 2014). The spread of signal crayfish has resulted in the widespread decline of native white-clawed crayfish (*Austropotamobius pallipes*). When tested in a similar experimental setting but with a shorter bridge length (Masefield, unpublished), a higher proportion (65%) of white-clawed crayfish left the water at least once compared to both invasive species tested in the present study (35.3% of red swamp and 46.6% of signal crayfish). Overall, however, white-clawed crayfish were found to leave the water less frequently than both invasive species. Further investigation of the tendency of native crayfish to leave the water is essential, given that conservation practices include the isolation of white-clawed crayfish in 'ark-sites', away from nearby waterbodies.

In terms of both red swamp and signal crayfish, the present study shows that terrestrial movement could be an equally important factor in the spread of both species, which are known to survive long periods out of water and in drought conditions (Cruz and Rebelo, 2007; Holdich

et al., 2014; Banha and Anastácio, 2014). The red swamp crayfish, however, is generally considered to be more adept at overland dispersal, tolerating long periods out of water (Piersanti et al., 2018) and extending survival on land by constructing burrows and inhabiting small puddles whilst also feeding on terrestrial vegetation (Grey and Jackson, 2012; Ramalho and Anastácio, 2015; Kouba et al., 2016; Souty-Grosset et al., 2016). In their native North American range, burrowing has not been recorded in signal crayfish and some studies show that this species is incapable of constructing burrows (Kouba et al., 2016), suggesting it is less adapted to terrestrial habitats. However, in Great Britain, invasive populations of signal crayfish frequently burrow in riverbanks and lakebeds (Holdich et al., 2014), which could explain their tendency to move overland in the present study, given the reduced risk of desiccation if they are able to construct burrows out of the water.

We have previously shown that ovigerous, non-ovigerous and juvenile signal crayfish from Britain also move out of water (Thomas et al., 2018) and in the present study we highlight that, at least in an experimental setting, signal crayfish are as likely to leave the water as red swamp crayfish, in the absence of competition or other stressors. As such, it is likely that signal crayfish leave the water and disperse overland more often than previously considered, which, coupled with its burrowing behaviour, could be a further factor contributing to its continuing invasion success in Great Britain (Holdich et al., 2014; Peay and Dunn, 2014). Given that the distance travelled overland in the present study was limited, however, further research should estimate the potential distance both species could travel overland in a field environment, which would be a useful avenue for future research to inform management and control practices.

Overall, this study shows that both species of invasive crayfish tested here move overland to a similar degree. Previous studies have shown that red swamp and signal crayfish both emerge on to land in response to dewatering of habitats, which can occur naturally or before biocide management treatments (Peay and Dunn, 2014; Ramalho and Anastácio, 2015) and such overland movement is likely to reduce the efficacy of control methods. Individuals that are prone to leaving the water will also be more likely to escape from commercial and private aquaculture ponds and enter nearby watercourses, increasing the risks of introduction and further range expansion (Holdich et al., 2014; Marques et al., 2015). Furthermore, terrestrial emigration allows crayfish to navigate in-stream barriers such as weirs or waterfalls (Kerby et al., 2005). The current study highlights the importance of alternate dispersal mechanisms which, despite not being widely reported, should be taken into consideration during management and population control practices of invasive species.

Declaration of Competing Interest

None.

Acknowledgement

This study was funded by the Coleg Cymraeg Cenedlaethol.

References

- Anastácio, P.M., Ferreira, M.P., Banha, F., Capinha, C., Rabaça, J.E., 2014. Waterbird-mediated passive dispersal is a viable process for crayfish (*Procambarus clarkii*). *Aquatic Ecol.* 48, 1–10. <https://doi.org/10.1007/s10452-013-9461-0>.
- Anastácio, P.M., Banha, F., Capinha, C., Bernardo, J.M., Costa, A.M., 2015. Indicators of movement and space use for two co-occurring invasive crayfish species. *Ecol. Indic.* 53, 171–181. <https://doi.org/10.1016/j.ecolind.2015.01.019>.
- Aquilioni, L., Ilhéu, M., Gherardi, F., 2005. Habitat use and dispersal of the invasive crayfish *Procambarus clarkii* in ephemeral water bodies of Portugal. *Mar. Freshw. Behav. Physiol.* 38, 225–236. <https://doi.org/10.1080/10236240500310195>.
- Banha, F., Anastácio, P.M., 2014. Desiccation survival capacities of two invasive crayfish species. *Knowl. Manag. Aquat. Ecosyst.* 413, 01. <https://doi.org/10.1051/kmae/2013084>.

- Barson, N.J., Cable, J., van Oosterhout, C., 2009. Population genetic analysis of micro-satellite variation of guppies (*Poecilia reticulata*) in Trinidad and Tobago: evidence for a dynamic source-sink metapopulation structure, founder events and population bottlenecks. *J. Evol. Biol.* 22, 485–497. <https://doi.org/10.1111/j.1420-9101.2008.01675.x>.
- Bowler, D.E., Benton, T.G., 2005. Causes and consequences of animal dispersal strategies: relating individual behaviour to spatial dynamics. *Biol. Rev.* 80, 205–225. <https://doi.org/10.1017/S1464793104006645>.
- Bubb, D.H., Lucas, M.C., Thom, T.J., Rycroft, P., 2002. The potential use of PIT telemetry for identifying and tracking crayfish in their natural environment. *Hydrobiologia* 483, 225–230. <https://doi.org/10.1023/A:1021352217332>.
- Bubb, D.H., Thom, T.J., Lucas, M.C., 2004. Movement and dispersal of the invasive signal crayfish *Pacifastacus leniusculus* in upland rivers. *Freshw. Biol.* 49, 357–368. <https://doi.org/10.1111/j.1365-2426.2003.01178.x>.
- Bubb, D.H., Thom, T.J., Lucas, M.C., 2006. Movement, dispersal and refuge use of co-occurring introduced and native crayfish. *Freshw. Biol.* 51, 1359–1368. <https://doi.org/10.1111/j.1365-2427.2006.01578.x>.
- Claussen, D.L., Hopper, R.A., Sanker, A.M., 2000. The effects of temperature, body size, and hydration state on the terrestrial locomotion of the crayfish *Orconectes rusticus*. *J. Crustacean Biol.* 20, 218–223. [https://doi.org/10.1651/0278-0372\(2000\)020\[0218:TEOTBS\]2.0.CO;2](https://doi.org/10.1651/0278-0372(2000)020[0218:TEOTBS]2.0.CO;2).
- Cote, J., Fogarty, S., Weinersmith, K., Brodin, T., Sih, A., 2010. Personality traits and dispersal tendency in the invasive mosquitofish (*Gambusia affinis*). *Proceedings of the Royal Society B* 277, 1571–1579. <https://doi.org/10.1098/rspb.2009.2128>.
- Cruz, M.J., Rebelo, R., 2007. Colonization of freshwater habitats by an introduced crayfish, *Procambarus clarkii*, in Southwest Iberian Peninsula. *Hydrobiologia* 575, 191–201. <https://doi.org/10.1007/s10750-006-0376-9>.
- Ellis, A., Jackson, M.C., Jennings, I., England, J., Phillips, R., 2012. Present distribution and future spread of Louisiana red swamp crayfish *Procambarus clarkii* (Crustacea, Decapoda, astacida, Cambaridae) in Britain: implications for conservation of native species and habitats. *Knowl. Manag. Aquat. Ecosyst.* 406. <https://doi.org/10.1051/kmae/2012022>.
- Enfjäll, K., Leimar, O., 2005. Density-dependent dispersal in the glanville fritillary, *Melitaea cinxia*. *Oikos* 108, 465–472. <https://doi.org/10.1111/j.0030-1299.2005.13261.x>.
- Gherardi, F., 2010. Invasive crayfish and freshwater fishes of the world. *Sci. Tech. Rev.* 29, 241–254. <https://doi.org/10.20506/rst.29.2.1973>.
- Gherardi, F., Barbaresi, S., 2000. Invasive crayfish: activity patterns of *Procambarus clarkii* in the rice fields of the Lower Guadalquivir (Spain). *Archiv für Hydrobiologie* 150, 153–168. <https://doi.org/10.1127/archiv-hydrobiol/150/2000/153>.
- Gherardi, F., Barbaresi, S., Salvi, G., 2000. Spatial and temporal patterns in the movement of *Procambarus clarkii*, an invasive crayfish. *Aquat. Sci.* 62, 179–193. <https://doi.org/10.1007/PL00001330>.
- Gherardi, F., Tricarico, E., Ilhéu, M., 2002. Movement and patterns of an invasive crayfish, *Procambarus clarkii*, in a temporary stream of southern Portugal. *Ethol. Ecol. Evol.* 14, 183–197. <https://doi.org/10.1080/08927014.2002.9522739>.
- Grey, J., Jackson, M.C., 2012. 'Leaves and eats shoots': direct terrestrial feeding can supplement invasive red swamp crayfish in times of need. *PLoS One* 7, e42575. <https://doi.org/10.1371/journal.pone.0042575>.
- Hanshew, B.A., Garcia, T.S., 2012. Invasion of the shelter snatchers: behavioural plasticity in invasive red swamp crayfish, *Procambarus clarkii*. *Freshw. Biol.* 57, 2285–2296. <https://doi.org/10.1111/fwb.12002>.
- Hill, A.M., Lodge, D.M., 1999. Replacement of resident crayfishes by an exotic crayfish: the roles of competition and predation. *Ecol. Appl.* 9, 678–690. [https://doi.org/10.1890/1051-0761\(1999\)009\[0678:RORCBA\]2.0.CO;2](https://doi.org/10.1890/1051-0761(1999)009[0678:RORCBA]2.0.CO;2).
- Holdich, D.M., Reynolds, J.D., Souty-Grosset, C., Sibley, P.J., 2009. A review of the ever-increasing threat to European crayfish from non-indigenous crayfish species. *Knowl. Manag. Aquat. Ecosyst.* 11, 394–395. <https://doi.org/10.1051/kmae/2009025>.
- Holdich, D.M., James, J., Jackson, C., Peay, S., 2014. The North American signal crayfish, with particular reference to its success as an invasive species in Great Britain. *Ethol. Ecol. Evol.* 26, 232–262. <https://doi.org/10.1080/03949370.2014.903380>.
- Hudina, S., Galie, N., Roessink, I., Hock, K., 2011. Competitive interactions between co-occurring invaders: identifying asymmetries between two invasive crayfish species. *Biol. Invasions* 13, 1791–1803. <https://doi.org/10.1007/s10530-010-9933-2>.
- Hudina, S., Hock, K., Žganec, K., 2014. The role of aggression in range expansion and biological invasions. *Curr. Zool.* 60, 401–409. <https://doi.org/10.1093/czoolo/60.3.401>.
- Hulme, P.E., Bacher, S., Kenis, M., Koltz, S., Kuhn, I., Minchin, D., Nentwig, W., Olenin, S., Panov, V., Pergl, J., Pysek, P., Roques, A., Sol, D., Solarz, W., Vila, M., 2008. Grasping at the routes of biological invasions: a framework for integrating pathways into policy. *J. Appl. Ecol.* 45, 403–414. <https://doi.org/10.1111/j.1365-2664.2007.01442.x>.
- Huner, J.V., Lindqvist, O.V., 1995. Physiological adaptations of freshwater crayfishes that permit successful aquacultural enterprises. *Integr. Comp. Biol.* 35, 12–19. <https://doi.org/10.1093/ich/35.1.12>.
- James, J., Slater, F., Cable, J., 2014. A. L. I. E. N. databases: addressing the lack in establishment of non-natives databases. *Crustaceana* 10, 1192–1199. <https://doi.org/10.1163/15685403-00003329>.
- Kerby, J.L., Riley, S.P.D., Kats, L.B., Wilson, P., 2005. Barriers and flow as limiting factors in the spread of an invasive crayfish (*Procambarus clarkii*) in Southern California streams. *Biol. Conserv.* 126, 402–409. <https://doi.org/10.1016/j.biocon.2005.06.020>.
- Kouba, A., Tíkal, J., Cisar, P., Veselý, L., Fořt, M., Přiborský, J., Patoka, J., Buřič, M., 2016. The significance of droughts for hyporheic dwellers: evidence from freshwater crayfish. *Sci. Rep.* 6, 26569. <https://doi.org/10.1038/srep26569>.
- Larson, E.R., Magoulick, D.D., Turner, C., Laycock, K.H., 2009. Disturbance and species displacement: different tolerances to stream drying and desiccation in a native and an invasive crayfish. *Freshw. Biol.* 54, 1899–1908. <https://doi.org/10.1111/j.1365-2427.2009.02243.x>.
- Lutz, C.G., Wolters, W.R., 1999. Mixed model estimation of genetic and environmental correlations in red swamp crayfish *Procambarus clarkii* (Girard). *Aquac. Res.* 30, 153–163. <https://doi.org/10.1046/j.1365-2109.1999.00249.x>.
- Marques, M., Banha, F., Águas, M., Anastácio, P., 2015. Environmental cues during overland dispersal by three freshwater invaders: *ericocheir sinensis*, *Pacifastacus leniusculus*, and *Procambarus clarkii* (Crustacea, Decapoda). *Hydrobiologia* 742, 81–93. <https://doi.org/10.1007/s10750-014-1968-4>.
- Peay, S., Holdich, D.M., Brickland, J., 2010. Risk assessments of non-indigenous crayfish in Great Britain. *Freshw. Crayfish* 17, 109–122.
- Peay, S., Dunn, A.M., 2014. The behavioural response of the invasive signal crayfish *Pacifastacus leniusculus* to experimental dewatering of burrows and its implications for eradication treatment and management of ponds with crayfish. *Ethol. Ecol. Evol.* 26, 277–298. <https://doi.org/10.1080/03949370.2014.903379>.
- Piersanti, S., Pallottini, M., Salerno, G., Goretti, E., Elia, A.C., Dörr, A.J.M., Reborá, M., 2018. Resistance to dehydration and positive hygrotaxi in the invasive crayfish *Procambarus clarkii*. *Knowl. Manag. Aquat. Ecosyst.* 419, 36. <https://doi.org/10.1051/kmae/2018024>.
- Puky, M., 2014. Invasive crayfish on land: *orconectes limosus* (Rafinesque, 1817) (Decapoda: cambaridae) crossed a terrestrial barrier to move from a side arm into the Danube River at Szermle, Hungary. *Acta Zoologica Bulgaria* 7, 143–146.
- R Core Team, 2017. R: a Language and Environment for Statistical Computing. URL: R Foundation for Statistical Computing, Vienna, Austria. <https://www.R-project.org/>.
- Ramalho, R.O., Anastácio, P.M., 2015. Factors inducing overland movement of invasive crayfish (*Procambarus clarkii*) in a ricefield habitat. *Hydrobiologia* 746, 135–146. <https://doi.org/10.1007/s10750-014-2052-9>.
- Rehage, J.S., Sih, A., 2004. Dispersal behavior, boldness, and the link to invasiveness: a comparison of four *Gambusia* species. *Biol. Invasions* 6, 379–391. <https://doi.org/10.1023/B:BINV.0000034618.93140.a5>.
- Roland, J., Keyghobadi, N., Fownes, S., 2000. Alpine *Parnassius* butterfly dispersal: effects of landscape and population size. *Ecology* 81, 1642–1653. <https://doi.org/10.2307/177313>.
- Shurin, J.B., Havel, J.E., 2002. Hydrologic connections and overland dispersal in an exotic freshwater crustacean. *Biol. Invasions* 4, 431–439. <https://doi.org/10.1023/A:1023692730400>.
- Souty-Grosset, C., Anastácio, P.M., Aquiloni, L., Banha, F., Choquer, J., Chucholl, C., Tricarico, E., 2016. The red swamp crayfish *Procambarus clarkii* in Europe: impacts on aquatic ecosystems and human well-being. *Limnologia* 58, 78–83. <https://doi.org/10.1016/j.limno.2016.03.003>.
- Strayer, D.L., 2010. Alien species in fresh waters: ecological effects, interactions with other stressors, and prospects for the future. *Freshw. Biol.* 55, 152–174. <https://doi.org/10.1111/j.1365-2427.2009.02380.x>.
- Strayer, D.L., Dudgeon, D., 2010. Freshwater biodiversity conservation: recent progress and future challenges. *J. North Am. Benthol. Soc.* 29, 344–358. <https://doi.org/10.1899/08-171.1>.
- Thomas, J.R., Fisher, J., Cable, J., Griffiths, S.W., 2018. Terrestrial dispersal of invasive signal crayfish during vulnerable life stages. *Behav. Processes* 157, 204–207. <https://doi.org/10.1016/j.beproc.2018.09.014>.
- Zuur, A.F., Ieno, E.N., Elphick, C.S., 2010. A protocol for data exploration to avoid common statistical problems. *Methods Ecol. Evol.* 1, 3–14. <https://doi.org/10.1111/j.2041-210X.2009>.